

## Research Article

**Cite this article:** Harrington TB (2020) Scotch broom (*Cytisus scoparius*) germination and growth responses to light: implications for logging debris retention after forest harvesting. *Weed Sci.* **68**: 517–526. doi: [10.1017/wsc.2020.53](https://doi.org/10.1017/wsc.2020.53)

Received: 7 April 2020

Revised: 11 June 2020

Accepted: 22 June 2020

First published online: 1 July 2020

**Associate Editor:**

Hilary A. Sandler, University of Massachusetts

**Keywords:**

Invasive plant species; light quality; red/far-red ratio; soil seedbank

**Author for correspondence:**

Timothy B. Harrington, Forestry Sciences Laboratory, 3625 93rd Avenue SW, Olympia, WA 98512-1101. (Email: [tim.harrington@usda.gov](mailto:tim.harrington@usda.gov))

# Scotch broom (*Cytisus scoparius*) germination and growth responses to light: implications for logging debris retention after forest harvesting

Timothy B. Harrington 

Research Forester, United States Department of Agriculture, Forest Service, Pacific Northwest Research Station, Olympia, WA, USA

**Abstract**

Scotch broom [*Cytisus scoparius* (L.) Link] is a large nonnative, leguminous shrub that threatens native plant communities by rapidly invading recently disturbed sites, competing vigorously for soil water and nutrients, and imparting soil legacy effects that inhibit native plants. In the Pacific Northwest, logging debris retention after forest harvesting prevents or slows *C. scoparius* invasions. A series of studies were conducted to determine potential mechanisms by which logging debris modifies the light environment to limit germination and growth of *C. scoparius*. In laboratory studies, seed germination did not vary significantly ( $P > 0.05$ ): (1) between presence and absence of light for several temperature regimes, (2) when exposed to red (660-nm wavelength) versus far-red (730-nm wavelength) light, and (3) across a range of red/far-red light (R/FR) ratios. These results indicate that modification of the light environment by logging debris or plant canopies has little or no influence on *C. scoparius* germination. In a study to simulate effects of variable mass of logging debris, “heavy” debris ( $2 \text{ kg m}^{-2}$ ) caused biologically relevant reductions in photosynthetic photon flux density (PPFD) and R/FR relative to conditions under “light” debris ( $1 \text{ kg m}^{-2}$ ). *Cytisus scoparius* germination did not differ significantly between simulated heavy and light debris; however, values of seedling root and shoot biomass under heavy debris were 16% and 71% of those observed under light debris, respectively. These results indicate that heavy debris limits biomass of *C. scoparius* seedlings, particularly roots, by reducing both PPFD and R/FR, which increases seedling vulnerability to summer drought or other stressors. Retention of heavy logging debris after forest harvesting has potential application on sites likely to be invaded by *C. scoparius*, as well as those sites with seedbanks containing *C. scoparius*.

**Introduction**

Scotch broom [*Cytisus scoparius* (L.) Link] is a large nonnative, leguminous shrub that threatens native plant communities by rapidly invading recently disturbed sites (Bossard and Rejmánek 1994; Peterson and Prasad 1998; Wearne and Morgan 2004), competing vigorously for soil water and nutrients (Carter et al. 2019; Watt et al. 2003b), and imparting soil legacy effects that inhibit native plant species (Grove et al. 2012, 2015). Native to western Europe, the species was introduced to Virginia in the early 1800s (Mack 2003) and to California in the 1850s (Mobley 1954), and it is currently found in 10 western and 19 eastern states of the United States (EDDMapS 2020). The species also has been introduced to Argentina, Australia, Bolivia, Brazil, Canada, Chile, China, India, Iran, Japan, New Zealand, and South Africa (CABI 2020; Peterson and Prasad 1998).

Although typically found in open sites (Peterson and Prasad 1998), *C. scoparius* can establish and survive in conditions of low light availability (10% of full sun; Williams 1981), such as a forest understory (Harrington 2007). In its native range, *C. scoparius* requires disturbance of associated vegetation in order to regenerate successfully, because newly germinated seedlings are relatively poor competitors (Paynter et al. 1998). In its introduced range, *C. scoparius* stands differ from those in its native range by having higher plant densities, multiple age classes, and less understory vegetation, all of which enable the stand to be self-replacing (Paynter et al. 2003).

*Cytisus scoparius* is a copious seed producer, with individual shrubs producing an average of 9,000 to 10,000 seeds  $\text{yr}^{-1}$  (Bossard and Rejmánek 1994; Smith and Harlen 1991). Ballistic dispersal transfers most seeds up to 1.5 m from the parent plant (Smith and Harlen 1991). Seeds have an impervious coat that causes innate dormancy (Young and Young 1992), but brief thermal shocks dramatically improve germination (Abdullah et al. 1989; Bossard 1993; Rivas et al. 2006; Smith and Harlen 1991) by melting or cracking the seed coat to make it permeable (Keeley 1987). Laboratory studies have identified that temperatures of 15 to 25 °C are favorable for germination of imbibed *C. scoparius* seeds; however, germination declines strongly for temperatures outside this range (Bossard 1993; Harrington 2009). Experimentally varying

© Weed Science Society of America, 2020. This is a work of the U.S. Government and is not subject to copyright protection in the United States.



the period of cold stratification from 0 to 60 d decreased the lag time for initiation of germination from 6 to 0 d but had no significant effect on cumulative 90-d germination (Harrington 2009).

Once buried, *C. scoparius* seeds are capable of germinating from depths  $\leq 8$  cm (Bossard 1993; Williams 1981). Seeds buried in the soil can have delayed germination for at least 4 yr (Bossard 1993; Smith and Harlen 1991), enabling development of seedbanks of  $<100$  to 21,000 seeds  $m^{-2}$  depending on stand maturity (Sheppard et al. 2002; Smith and Harlen 1991). In the absence of seed rain, *C. scoparius* seedbanks are estimated to decline 40% to 50% each year from germination and seed loss (Paynter et al. 1998; Sheppard et al. 2002).

As a leguminous species, *C. scoparius* can fix over 100 kg  $ha^{-1} yr^{-1}$  of N in its aboveground biomass and return 17 kg N  $ha^{-1} yr^{-1}$  to the soil as leaf litter (Watt et al. 2003a). Benefits of these nitrogen amendments to associated trees, such as radiata pine [*Pinus radiata* D. Don] in Australia, are minimal to nonexistent, because 3 yr of *C. scoparius* competition can cause a 96% reduction in tree biomass relative to trees grown in the absence of *C. scoparius* (Watt et al. 2003b). *Cytisus scoparius* can have multiple soil legacy effects, including accumulation of soil carbon and nitrogen and depletion of soil phosphorus and potassium (Caldwell 2006; Haubensak and Parker 2004; Slesak et al. 2016a), which can favor nonnative grasses (Carter et al. 2019; Grove et al. 2015; Shaben and Myers 2010). Reduced biomass of coast Douglas-fir (*Pseudotsuga menziesii* (Mirb.) Franco var. *menziesii*) seedlings grown in *C. scoparius*-invaded soils was attributed to allelopathic effects (Grove et al. 2012).

Controlling *C. scoparius* can be expensive and challenging. Direct and indirect annual costs for treatment expenses and lost timber revenues have been estimated to be \$40 million for Oregon (Hulting et al. 2008) and \$143 million for Washington (CAI 2017). Tactics for controlling or preventing invasions of *C. scoparius* in forest and rangelands include cutting, prescribed fire, herbicide treatments, and retention of logging debris after forest harvesting (Bossard and Rejmánek 1994; DiTomaso et al. 2006; Oneto et al. 2010; Peter and Harrington 2018).

Because *C. scoparius* has a persistent seedbank (i.e., a longevity of at least 5 yr, and likely much longer; Bossard 2000), PRE control is essential to managing forest sites already invaded by this species. At operational rates, metsulfuron and hexazinone herbicides provided 75% and 90% control of emerging *C. scoparius* seedlings, respectively; whereas, sulfometuron only reduced biomass of emerging seedlings but did not affect their survival (Ketchum and Rose 2003). Sulfometuron and metsulfuron have been shown to limit root development of *C. scoparius* seedlings and thereby increase their susceptibility to soil drought (Harrington 2009). In PRE applications, maximum labeled rates of the synthetic auxin herbicides aminopyralid, aminocyclopyrachlor, and clopyralid killed an average of 85% of emerging *C. scoparius* seedlings; however, even rates as low as 25% of the maximum killed more than 60% of seedlings (Harrington 2014).

In western Washington USA, retaining logging debris after forest harvesting (i.e., stem-only harvesting; 2 to 2.5 kg  $m^{-2}$  of debris) strongly reduced density and cover of *C. scoparius* seedlings relative to conventional whole-tree harvesting (1 to 1.5 kg  $m^{-2}$  of debris) (Harrington and Schoenholtz 2010; Peter and Harrington 2018). Logging debris retention also improved survival and growth of planted seedlings of coast Douglas-fir (Harrington and Schoenholtz 2010; Harrington et al. 2018), and it fostered the development of a native plant community (Peter and Harrington 2018).

This report describes a series of studies that were conducted to determine some of the mechanisms by which logging debris modifies the light environment to limit germination and growth of *C. scoparius*. Two components of the light environment relevant to plants include light quantity (i.e., photosynthetic photon flux density [PPFD]) and light quality (i.e., spectral distribution of PPFD by active wavelengths, such as red and far-red light). The author is not aware of any studies evaluating the effects of light quantity or quality on *C. scoparius* germination. An improved understanding of plant responses to light is needed if the light environment of cropping systems is to be manipulated successfully to improve weed management (Holt 1995). Laboratory studies addressed the following three research questions. (1) Is light required for *C. scoparius* seed germination? (2) Do *C. scoparius* germination and growth differ when exposed to red versus far-red light? (3) Does *C. scoparius* germination vary with exposure to different red/far-red light (R/FR) ratios? A field study addressed a fourth study question: (4) How does variation in mass of simulated logging debris influence microclimate and germination and growth of *C. scoparius*?

## Materials and Methods

### Seed Collection, Storage, and Quality Control

In 2012, a bulk collection of seeds from multiple *C. scoparius* plants was conducted near Matlock, WA (47.215°N, 123.442°W) for the simulated logging debris effects study. In that study, seed mortality from biocontrol insects accounted for almost half of the test population, as evidenced by the collapsing, liquification, and rotting of seeds hollowed out by insect larvae. To understand the biological mechanisms driving *C. scoparius* seed germination responses to its light environment, an experimental approach was needed in subsequent studies to account for genetic variation in seed mortality from biocontrol insects, such as the use of experimental blocking by parent-plant seed source. Beginning in 2013, seeds were collected and bagged separately from individual parent plants located near Shelton, WA (47.302°N, 123.200°W) and Tumwater, WA (46.986°N, 122.891°W). Each year a different set of parent plants was identified for seed collection, and selected plants were spaced at least 3 m apart. As soon as they began to dehisce in mid-July, seed pods were collected and air-dried in paper bags to stimulate their opening. Extracted seeds were stored in sealed glass containers at  $-13$  C. Seeds from the previous growing season were used in each study, and during counting they were visually inspected to eliminate those that were visibly damaged, unusually small, or immature (i.e., green). The actual number of parent-plant seed sources available for a given study (i.e., the number of experimental blocks) depended on variation in annual seed production and the estimated number of potentially viable seeds found after visual inspection.

### Germination and Growth Responses to Light

#### Effects of Light and Temperature

To determine whether light is required for *C. scoparius* germination, in December 2018 two sets of 50 seeds from each of nine parent plants were placed on moistened paper trays in covered petri dishes that were subjected to a daily dark/light (14/10 h) temperature regime of 10/15 C. Fluorescent lighting in the germination chamber provided an average of 65  $\mu\text{mole m}^{-2} \text{s}^{-1}$  of PPFD, as measured by a quantum sensor (LI-190SA quantum sensor, Li-Cor, 4647 Superior Street, Lincoln, NE 68504, USA). The R/FR

of the fluorescent lighting averaged 4.0 based on readings from a LightScout red/far-red meter (Spectrum Technologies, 3600 Thayer Court, Aurora, IL 60504, USA). Green headlamps (Apollo Horticulture 17 Watt LED High Intensity Green Light Headlamp, <https://www.apollohorticulture.com>) were used while handling seeds to prevent light from influencing their subsequent responses to treatment. Edges of petri dishes were wrapped in parafilm to prevent moisture loss (Parafilm M, Heathrow Scientific, 620 Lakeview Parkway, Vernon Hills, IL 60061, USA). In one set, petri dishes were also wrapped in aluminum foil to eliminate incoming light. A germinant was defined as any seed visibly showing the white hypocotyl of a *C. scoparius* embryo. After 20 d, germination percentages were calculated for each petri dish using the following equation:

$$G = 100 \times L \div (R + L) \quad [1]$$

where  $G$  is cumulative germination (%),  $L$  is the number of living germinants, and  $R$  is the number of remaining intact seeds presumed to be alive. The equation provides a more accurate estimate of seed germination responses to experimental treatments, because it includes only those seeds having the biological potential to germinate. Seed mortality percentages were calculated for each petri dish using the following equation:

$$M = 100 \times D \div (R + L + D) \quad [2]$$

where  $M$  is cumulative seed mortality (%),  $D$  is the number of dead seeds (i.e., those found to be rotten, liquified, or empty), and all other variables are defined as for Equation 1. The experiment was repeated for daily dark/light temperature regimes of 15/20 C and 20/25 C in February and March 2019, respectively.

#### Effects of Red versus Far-Red Light

To determine whether *C. scoparius* germination differs when exposed to red versus far-red light, two germination chambers were outfitted with either four red (660-nm wavelength) or four far-red LED bulbs (730-nm wavelength) (powerPAR 15 watt LED grow light, Hydrofarm, 2249 S. McDowell Ext., Petaluma, CA 94954, USA). Measurements of total irradiance taken with a Jaz® Combo spectrometer (Ocean Optics, 830 Douglas Avenue, Dunedin, FL 34698, USA) indicated that energy outputs differed about 22% between the red (1,227 W m<sup>-2</sup>) and far-red (1,007 W m<sup>-2</sup>) germination chambers. Each germination chamber was programmed for a daily dark/light (14/10 h) temperature regime of 15/20 C. In October 2017, two sets of 50 seeds from each of 12 parent plants were placed on moistened paper trays in covered petri dishes; one set was placed in the red germination chamber and the other set was placed in the far-red chamber. To confirm the experimental validity of the red and far-red light treatments, three additional petri dishes were included in each germination chamber that contained seeds of Grand Rapids leaf lettuce (*Lactuca sativa* L.)—a variety known to exhibit inhibition of germination in response to far-red light (Borthwick et al. 1952). Over a 20-d period, germinants were counted and removed every 1 to 2 d. Immediately before each count, the petri dishes were rotated to remove potential location effects, and paper trays were remoistened with 1 to 2 ml of distilled water. Dead seeds (as evidenced by blackening or liquification) also were counted and removed from each petri dish. Cumulative germination and seed mortality percentages were calculated per petri dish with Equations 1 and 2, respectively. The study was repeated in May and June 2019 with seeds from eight and six parent plants, respectively (i.e., in these later trials,

collections from some parent plants did not have adequate numbers of potentially viable seeds to be included in the experiment). Upon completion of the May and June 2019 studies, remaining *Lactuca sativa* seeds in the far-red germination chamber were moved to the red chamber to confirm the reversibility of germination inhibition (Borthwick et al. 1952).

To compare growth of *C. scoparius* seedlings when exposed to red versus far-red light, in August 2017, a total of 20 plastic boxes of dimensions 17 cm by 12 cm by 6 cm with drain holes were filled with potting soil (Sunshine® Mix #1, Sun Gro Horticulture, 770 Silver Street, Agawam, MA 01001, USA). Potting soil was pre-wetted to ensure uniformity of moisture within each box. Each box was labeled to hold seeds from two parent plants, one in each half of the box. Two sets of 50 seeds each were counted out for each of the 10 parent plants. Working within the lighting of the germination chambers only, seed sets were placed on the soil surface of the appropriately labeled end of each box. Seeds were pressed into the soil to ensure good contact, but they were not fully buried. Every 2 d, the soil boxes were rotated to remove potential location effects, and approximately 10 ml of distilled water was added to each box to maintain moisture content at a moderate level. After 40 d, seedlings within each soil box were counted and measured for height (nearest 0.1 cm). Seedlings were then carefully excavated, their roots washed, and their roots and shoots composited into separate paper envelopes. Root and shoot biomass values (nearest mg) were determined after drying at 65 C to a constant weight. The study was repeated in October 2017 with seeds from nine parent plants.

#### Effects of Red/Far-Red Ratio

To determine whether *C. scoparius* germination differs when exposed to various ratios of red/far-red light, a total of six trials were conducted from July to October 2019 within a germination chamber outfitted with one of six combinations of red and far-red LED light bulbs: 0/4, 1/4, 2/3, 3/2, 4/1, or 5/0, respectively. R/FR was measured for each lighting combination with a LightScout red/far-red meter, resulting in the following values: <0.01, 0.4, 0.9, 1.6, 3.6, and 31.1. For each trial, 50 seeds from each of 15 parent plants were placed in moistened petri dishes and subjected to a daily dark/light (14/10 h) temperature regime of 15/20 C for 20 d. Every 1 to 2 d, 1 to 2 ml of distilled water was added to each petri dish, and germinated seeds and dead seeds were counted separately and removed. One additional petri dish was included in each trial that contained 50 seeds of *Lactuca sativa* to confirm the experimental validity of the R/FR treatments. Cumulative germination and seed mortality percentages were calculated per petri dish with Equations 1 and 2, respectively.

#### Simulated Logging Debris Effects on Microclimate, Germination, and Growth

To simulate the effects of various masses of logging debris on microclimate and germination and growth of *C. scoparius*, 15 frames were constructed of 1.9-cm-diameter PVC and 1.3-cm mesh hardware cloth in summer 2012. Three tables (i.e., experimental blocks) were placed parallel to each other in a nearby open area outside the Forestry Sciences Laboratory, Olympia, WA (46.952°N, 122.959°W), and five frames were attached to each table. Each frame had a 60 cm by 60 cm screen platform for suspending logging debris 20 cm above the table surface. The tabletops and outside vertical surfaces of each frame were wrapped in black

plastic to limit light to only that entering through the screen platform. In August 2012, fresh branches of coast Douglas-fir  $\leq 1.3$  cm in diameter (i.e., simulated logging debris) were collected from nearby trees. Five evenly spaced masses of logging debris were selected to encompass the range of 0 to 2.78 kg m<sup>-2</sup> and to include levels designated previously as typical of light (1 kg m<sup>-2</sup>) and heavy (2 kg m<sup>-2</sup>) retention after forest harvesting (Harrington et al. 2018). One mass of logging debris was randomly assigned to each frame on a given table. R/FR was measured immediately after application of the fresh green debris via a Jaz® Combo spectrometer. Measurement of R/FR was repeated in October 2012 after desiccation and browning of the debris. Both measurements of R/FR were taken on cloudless days near solar noon.

In late April 2013, a 90-d experiment was initiated with application of fresh samples of logging debris to each frame according to the 2012 treatment assignments. Air temperature was recorded every 2 h via an iButton sensor (model DS1921G, Maxim Integrated, 160 Rio Robles, San Jose, CA 95134, USA) suspended within a small plastic bag approximately 10 cm below the screen platform near the center of each debris frame. PPFD was measured within each frame with an AccuPAR LP-80 ceptometer (Meter Group, 2365 NE Hopkins Court, Pullman, WA 99163, USA). Measurements of PPFD were taken on cloudless days near solar noon when the debris was green (May) and again after the debris had desiccated and turned brown (July). Two readings, taken diagonally under the suspended logging debris of each frame, were averaged and then expressed as a percentage of an open-sky reading taken facing south.

To quantify the relationship of *C. scoparius* germination to logging debris mass, approximately 150 seeds (i.e., 1.5 g) from a bulk collection (i.e., not separated by individual parent plants) were placed on a moistened paper tray in each of 15 covered petri dishes; each dish was enclosed within a sealed plastic bag and placed under the suspended logging debris of a given frame. Every 1 to 2 d, the paper tray within each petri dish was remoistened with 1 to 2 ml of distilled water, and germinated seeds and dead seeds were counted separately and removed. On day 90, cumulative germination and seed mortality percentages were calculated per petri dish with Equations 1 and 2, respectively.

To evaluate the relationship of *C. scoparius* seedling biomass to logging debris mass, approximately 150 seeds (i.e., 1.5 g) from a bulk collection (i.e., not separated by individual parent plants) were sown into each of 30 plastic soil boxes using the methods described earlier. Soil for 15 boxes came from a site near Matlock, WA (47.206°N, 123.442°W)—a very gravelly loamy sand of the Grove series (sandy-skeletal, mixed, mesic Dystric Xerorthents) formed in glacial outwash (USDA-NRCS 2019). Soil for the remaining 15 boxes came from a site near Molalla, OR (45.196°N, 122.285°W)—a cobbly loam of the Kinney series (fine-loamy, isotic, mesic Andic Humudepts) (USDA-NRCS 2019). Before seeds were sown, each soil was sieved to remove coarse fragments greater than 2 mm, each box was filled with 1 kg of soil, and distilled water was added to achieve field capacity. One Matlock soil box and one Molalla soil box were placed under the suspended logging debris of each frame. At weekly intervals, distilled water was added to each soil box to restore field capacity. Soil boxes from one experimental block were weighed immediately before watering to estimate evaporative water loss. On day 90, height (nearest 0.1 cm) of the five tallest seedlings per soil box were measured. Seedlings within each soil box were counted and carefully excavated, their roots were washed, and shoots and roots were

composited into separate paper envelopes. Root and shoot biomass values (nearest mg) were determined after drying at 65 C to a constant weight.

### Statistical Analyses

All statistical analyses were conducted in SAS (SAS System for Windows v. 9.4, SAS Institute, Cary, NC 27513, USA) with a significance level of  $\alpha = 0.05$ . Most experiments were repeated one or two times, and statistical analyses were conducted separately for each trial. The experimental design of all laboratory experiments was a randomized complete block with blocks assigned to parent-plant seed source to control for genetic variation in seed mortality from biocontrol insects; experimental units were individual petri dishes or soil box halves. An angular transformation was applied to proportionate germination values, and a logarithmic transformation was applied to seedling height and biomass values to homogenize the residual variation before ANOVA (Sokal and Rohlf 1981). Residuals from ANOVA and regression were plotted against predicted values to confirm the homogeneity of the residual variances. All reported values of the coefficient of determination were adjusted for degrees of freedom (i.e., adjusted R<sup>2</sup>).

### Effects of Light and Temperature

Final cumulative germination data were subjected to ANOVA in SAS PROC MIXED to test the significance of the fixed factors, light (presence vs. absence), temperature regime, and their interaction, after adjusting for random effects of parent plant (i.e., blocks). Orthogonal polynomial contrasts were included in the ANOVA to test for linear and quadratic effects of temperature regime. A similar analysis was conducted for the final cumulative seed mortality data.

### Effects of Red versus Far-Red Light

Incremental values of cumulative germination of *C. scoparius* or *Lactuca sativa* were subjected to repeated-measures ANOVA in SAS PROC MIXED to test the significance of the fixed factor, light treatment (i.e., red vs. far-red light), day of trial, and their interaction, after adjusting for random effects of parent plant. If a treatment-by-day interaction was detected, slicing was used in SAS to identify individual days in which differences existed between treatments. A similar analysis was conducted for the cumulative seed mortality data. Mean values of seedling height and root and shoot biomass for each trial were subjected to ANOVA in SAS PROC MIXED to test the significance of the light treatment after adjusting for random effects of parent plant.

### Effects of Red/Far-Red Ratio

Final cumulative germination data for *C. scoparius* were subjected to ANOVA in SAS PROC MIXED to test the significance of the fixed factor, R/FR (logarithmically transformed to provide approximately equal spacing), after adjusting for random effects of parent plant. Orthogonal polynomial contrasts were included in the ANOVA to test for linear and quadratic effects of R/FR. Polynomial regression was used to characterize the germination responses of *Lactuca sativa* to R/FR. A similar analysis was conducted for the final cumulative seed mortality data.

### Simulated Logging Debris Effects on Microclimate, Germination, and Growth

Values of mean daily air temperature by month, PPFD, R/FR, cumulative germination, cumulative seed mortality, seedling

height, and root and shoot biomass were subjected to ANOVA in SAS PROC MIXED to test the significance of the fixed factor, debris mass, after adjusting for random effects of blocks. Orthogonal polynomial contrasts were included in the ANOVA to test for linear, quadratic, and cubic effects of debris mass. Results of the contrasts were used to fit appropriate polynomial or nonlinear models via SAS PROC REG or PROC NLIN, respectively, to illustrate each variable's response to debris mass. A negative exponential model was fit to the relationship of soil box weight loss versus logging debris mass. The coefficient of determination for nonlinear models was estimated from the linear regression of observed versus predicted values. For air temperature, analysis and reporting were focused on data from May, because the range of mean daily values for that month (12.8 to 14.3 C) was biologically relevant to potential temperature constraints on seed germination (Bossard 1993; Harrington 2009). For both PPFD and R/FR, models with separate regression coefficients for green (i.e., fresh) and brown (i.e., desiccated) debris were fit to the pooled data, and the extra sums-of-squares approach was used to select the simplest model for each variable (Neter et al. 1989). For soil box weight loss, seedling height, and root and shoot biomass, differences due to soil type were tested similarly using the extra sums-of-squares approach.

#### Predicting Temperature Limitations to Germination

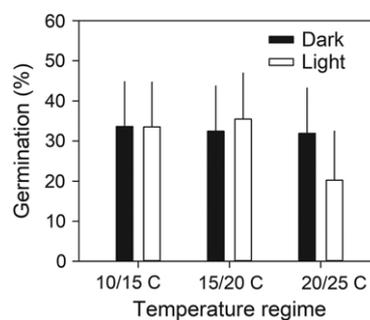
To illustrate potential temperature limitations to *C. scoparius* germination as a result of logging debris retention, germination in the controls (i.e., absent thermal shock treatments) for temperatures less than 25 C were interpolated from Figures 2 and 3 of Bossard (1993) for a 24-d period to match the experimental duration used by Harrington (2009). Germination data for the same temperature range from Figure 1 of Harrington (2009) were adjusted to remove effects of a 45-d cold-stratification period by multiplying observed germination by 0.63 based on the average values shown in Figure 2 of Harrington (2009) for day 24. Differences in daily photoperiods between data sources (Bossard [1993]: 15-h light/9-h dark; Harrington [2009]: 10-h light/14-h dark) were assumed to be inconsequential. Temperatures for each regime in Bossard (1993) were held constant, but the light and dark temperatures for each regime in Harrington (2009) differed by 5 C. Linear regression was used to characterize the relationship of *C. scoparius* germination to temperature for the combined data from Bossard (1993) and Harrington (2009). Differences in the relationship due to data source were tested via SAS PROC REG using the extra sums-of-squares approach (Neter et al. 1989). Using the upper (i.e., lighted) temperature value for each regime from Harrington (2009) in the regression resulted in a relationship that closely matched that found for the Bossard (1993) data.

## Results and Discussion

### Germination and Growth Responses to Light

#### Effects of Light and Temperature

Cumulative germination of *C. scoparius* did not differ significantly among levels of light ( $P = 0.40$ ), temperature ( $P = 0.15$ ), or their interaction ( $P = 0.22$ ) (Figure 1). Cumulative seed mortality from biocontrol insects averaged 31% and did not differ significantly among levels of light ( $P = 0.08$ ), temperature ( $P = 0.90$ ), or their interaction ( $P = 0.91$ ). However, stem lengths of seedlings germinating in darkness were more than three times those of seedlings



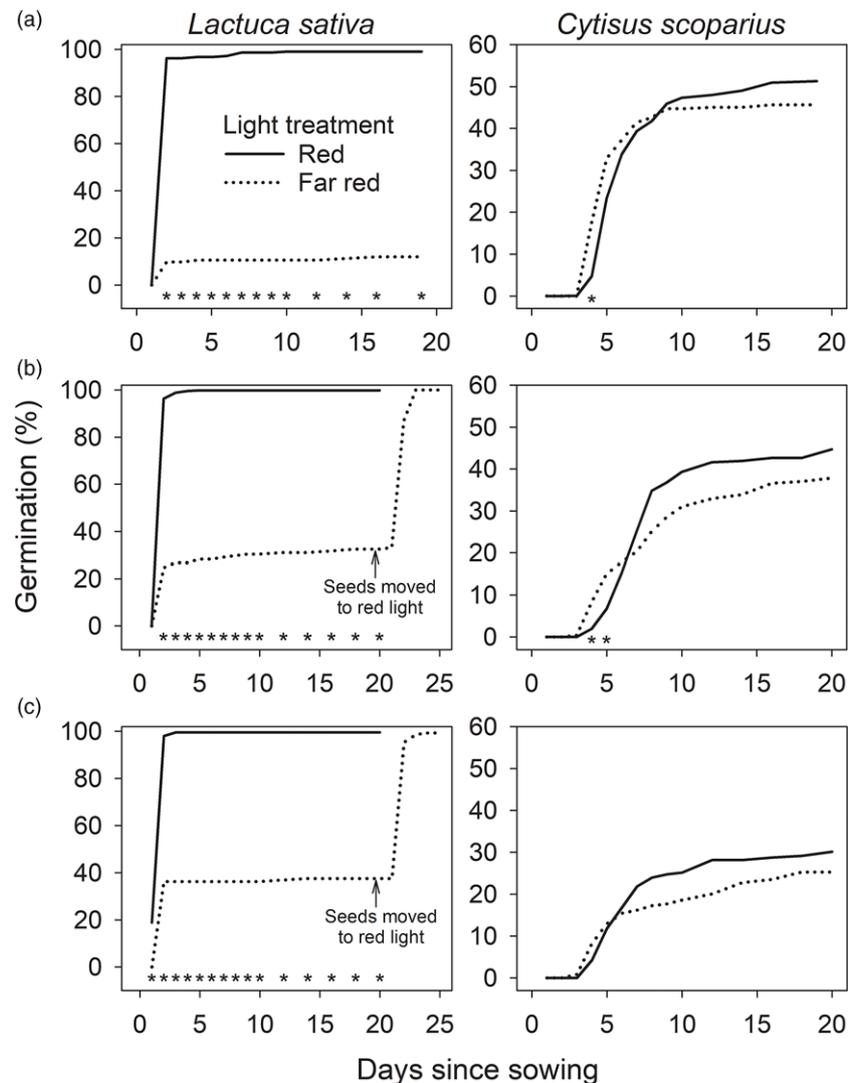
**Figure 1.** *Cytisus scoparius* germination ( $\pm$ SE) as affected by light and temperature in a laboratory study in Olympia, WA, during December 2018 to March 2019.

germinating in light. Reduced germination in the lighted 20/25 C regime was associated with partial drying in some petri dishes, but it did not change overall results of the study. This study confirms that *C. scoparius* is not photoblastic (i.e., germination is neither negatively nor positively responsive to light).

#### Effects of Red versus Far-Red Light

During the first 5 d of each trial, cumulative germination of *C. scoparius* was greater under far-red light than under red light – a difference that was statistically significant in the first two trials ( $P \leq 0.039$ ) (Figure 2). This brief stimulation of germination by far-red light could confer a slight, short-term advantage for *C. scoparius* establishment under green logging debris or plant canopies. Later in each trial, *C. scoparius* germination rates were greater under red light. At the end of each trial, cumulative germination of *C. scoparius* under red light averaged 5% to 7% greater than under far-red light, although differences were not statistically significant ( $P \geq 0.29$ ). Cumulative seed mortality from biocontrol insects averaged 38% for the three trials and did not differ significantly between red and far-red light ( $P \geq 0.493$ ). Cumulative germination of *Lactuca sativa* under far-red light was 32% to 38% of that observed under red light; however, moving the ungerminated seeds into the red-light germination chamber resulted in 100% germination (Figure 2). These results indicate that *C. scoparius* germination is not subject to phytochrome regulation, whereas *Lactuca sativa* germination is clearly phytochrome regulated. Gorse (*Ulex europaeus* L.), an invasive leguminous shrub with life history traits similar to those of *C. scoparius*, also has seed germination that is not photoblastic (Ivens 1983) or subject to phytochrome regulation (JA Zabkiewicz, personal communication, cited in Rolston and Talbot 1980).

In the two experimental trials of the biomass study, seedling numbers ranged from 0 to 7 per soil box. Height of *C. scoparius* seedlings from each of the trials did not differ significantly between red and far-red light ( $P \geq 0.110$ ; Table 1). Root biomass of *C. scoparius* seedlings grown in red light averaged 2.4 and 6.8 times greater in Trials 1 and 2, respectively, than that of seedlings grown in far-red light ( $P \leq 0.046$ ; Table 1). Shoot biomass also was greater in red light than in far-red light, but differences were significant only in Trial 2 ( $P = 0.017$ ). In a meta-analysis of biomass responses of terrestrial plants to environmental stress, allocation to stem increased but allocation to leaves decreased as R/FR decreased below 0.9, but there was no clear response for roots (Poorter et al. 2012). In the current study, perhaps reduced biomass allocation to roots in the far-red chamber can be attributed to lower radiant energy ( $1,007 \text{ W m}^{-2}$ ) than in the red chamber ( $1,227 \text{ W m}^{-2}$ ).



**Figure 2.** Germination of *Lactuca sativa* (Grand Rapids variety) (left column) and *Cytisus scoparius* (right column) when exposed to red versus far-red light in: (A) October 2017, (B) May 2019, and (C) June 2019 in laboratory studies in Olympia, WA. For each species, asterisks indicate dates in which cumulative germination differed significantly ( $P \leq 0.05$ ) between red and far-red light.

#### Effects of Red/Far-Red Ratio

Cumulative germination of *C. scoparius* did not vary significantly among R/FR values ( $P = 0.397$ ), averaging 62% (Figure 3). Linear and quadratic effects of R/FR also were nonsignificant for *C. scoparius* ( $P \geq 0.08$ ). Cumulative seed mortality from biocontrol insects averaged 67% and did not differ significantly among values of R/FR ( $P = 0.653$ ). Cumulative germination of *Lactuca sativa* had a quadratic relationship with  $\log_{10}(\text{R/FR})$  ( $R^2 = 0.99$ ) because of the contrasting response between R/FR = 0.01 (32% germination) and R/FR  $\geq 0.38$  (88% to 100% germination) (Figure 3).

#### Simulated Logging Debris Effects on Microclimate, Germination, and Growth

Mean daily air temperature in May varied significantly among debris masses ( $P = 0.001$ ). Linear, quadratic, and cubic effects of debris mass all were significant ( $P \leq 0.02$ ). A cubic regression model provided the best fit to the relationship of air temperature to debris mass ( $R^2 = 0.68$ ; Table 2). As simulated debris mass increased from 1 kg m<sup>-2</sup> ("light debris") to 2 kg m<sup>-2</sup> ("heavy

debris"), predicted air temperature in May decreased 0.5 C from 13.6 to 13.1 C (Figure 4A). Harrington et al. (2018) observed reductions in soil temperature at 5-cm depth of 0.5 to 1.5 C for heavy versus light debris during late spring and summer.

Although weekly soil box weight loss from evaporation exhibited high levels of variation, it had a significant negative exponential relationship with logging debris mass that did not differ between soil sources ( $R^2 = 0.15$ ; Table 2). As simulated debris mass increased from 1 to 2 kg m<sup>-2</sup>, predicted soil box weight loss decreased from 14% to 10% (Figure 4B). Decreases in soil water loss due to mulching effects of logging debris have been observed in previous research (Harrington et al. 2018; O'Connell et al. 2004; Roberts et al. 2005), and generally they have been attributed to reductions in soil temperature and evaporation (Devine and Harrington 2007).

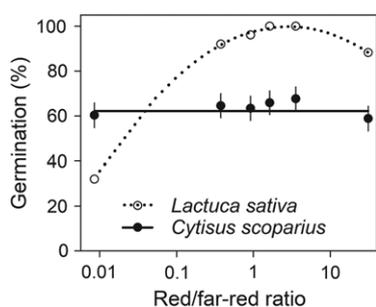
PPFD differed significantly among debris masses for both green and brown debris ( $P < 0.001$ ). For both types of debris, linear and quadratic effects of debris mass were significant ( $P \leq 0.003$ ); however, a negative exponential regression model provided the best fit for these data ( $R^2 = 0.99$ ; Table 2). Attenuation of PPFD

**Table 1.** Average height and root and shoot biomass of *Cytisus scoparius* seedlings (SE in parentheses) when exposed to red versus far-red light in laboratory studies in Olympia, WA, during August (Trial 1) and October (Trial 2) 2017.<sup>a</sup>

Trial	Variable	Treatment		df <sup>b</sup>		Prob. > F
		Red light	Far-red light	Num	Den	
1	Height (cm)	2.2 (0.3)	2.9 (0.3)	1	7	0.110
	Root biomass (mg)	15.3 (4.0)	6.3 (1.7)	1	7	0.046
	Shoot biomass (mg)	12.2 (1.4)	9.8 (1.1)	1	7	0.212
2	Height (cm)	3.4 (0.4)	3.8 (0.3)	1	3	0.509
	Root biomass (mg)	5.4 (1.6)	0.8 (0.2)	1	3	0.013
	Shoot biomass (mg)	14.5 (2.6)	4.9 (0.7)	1	3	0.017

<sup>a</sup>Seedlings were grown for 40 d in a standard potting soil.

<sup>b</sup>df for the numerator (Num) and denominator (Den) of the F-test from ANOVA.



**Figure 3.** Germination of *Lactuca sativa* (Grand Rapids variety) and *Cytisus scoparius* ( $\pm$ SE) when exposed to various red/far-red light ratios (R/FR) in a laboratory study in Olympia, WA, during July to October 2019. Trend lines are from polynomial regression (*Lactuca sativa*:  $Y = 97.7 + 9.67(X) - 10.7(X^2)$ , where  $X = \log_{10}(R/FR)$ ,  $P < 0.001$ ,  $R^2 = 0.99$ ,  $n = 6$ ; *C. scoparius*:  $Y = 62.2$ ,  $P = 0.920$ ,  $n = 90$ ).

occurred more rapidly with increasing debris mass for green debris than for brown debris (Figure 4C). For example, predicted PPFD values for light and heavy debris are 11% and 2% of full sun for green debris, whereas they are 30% and 12% for brown debris, respectively. Greater attenuation of PPFD under fresh debris can be attributed both to its green color, which is effective at absorbing blue and red light, and its greater surface area.

R/FR differed significantly among debris masses for both green and brown debris ( $P \leq 0.002$ ). Linear effects of debris mass were significant for both types of debris ( $P < 0.001$ ); however, as found for PPFD, a negative exponential regression model provided the best fit for these data ( $R^2 = 0.87$ ; Table 2). R/FR declined rapidly with increasing debris mass for green debris, whereas it declined slowly for brown debris (Figure 4D). For example, predicted R/FR values for light and heavy debris are 0.5 and 0.1 for green debris, whereas they are 1.2 and 1.1 for brown debris, respectively. These model predictions suggest that only green debris can alter R/FR sufficiently to elicit a biological response in a neighboring plant; brown debris transmits enough red light to cancel potential biological effects of far-red light. Bliss and Smith (1985) found that fresh litter of English oak (*Quercus robur* L.) attenuated red light more than far-red light, because the overlapping, pigmented material acted as a filter to screen out specific wavelengths, similar to that found for green Douglas-fir logging debris in the current study. In contrast, fresh litter of Corsican pine (*Pinus nigra*

Arnold var. *maritima*) had a neutral effect on spectral transmission, similar to that found for brown Douglas-fir logging debris, because ambient light penetrated the loosely packed material and the pine needles themselves transmitted very little light because of their thickness (1 mm) (Bliss and Smith 1985).

Cumulative germination of *C. scoparius* did not differ significantly among debris masses ( $P = 0.199$ ). Linear, quadratic, and cubic effects of debris mass also were not significant ( $P > 0.260$ ). Mean germination for the 90-d trial was 32% (Table 2; Figure 5A). Although *C. scoparius* germination was variable in this study, absence of responses to varying debris mass provides additional support for the species' non-photoblastic response. Cumulative seed mortality from biocontrol insects averaged 48% and did not differ significantly among debris masses ( $P \geq 0.818$ ).

Height of *C. scoparius* seedlings differed significantly among debris masses ( $P < 0.001$ ) and soil sources ( $P = 0.002$ ), but it did not differ significantly with the interaction of debris mass and soil source ( $P = 0.242$ ). Only quadratic effects of debris mass were significant for height ( $P = 0.001$ ). A quadratic regression model provided the best fit to the height responses (Table 2). Based on the regression model, *C. scoparius* height peaked in each soil type at a debris mass of  $1.64 \text{ kg m}^{-2}$  (Figure 5B); R/FR under green debris of this mass is predicted to be 0.2—a value that has been definitively shown to stimulate the classic shade-avoidance response of increased stem elongation (Poorter et al. 2012; Smith 1982). Reduced height of *C. scoparius* seedlings grown under the highest debris mass ( $2.78 \text{ kg m}^{-2}$ ) is likely indicative of severe limitations in seedling productivity from restricted light availability (i.e., predicted PPFD  $< 6\%$  of full sun; see following discussion of biomass responses). Seedling height averaged 2 cm greater when grown on soil from Molalla versus that from Matlock, probably because of the superior quality of Molalla soil (Slesak et al. 2016b).

Root and shoot biomass of *C. scoparius* seedlings both differed significantly among debris masses ( $P \leq 0.001$ ), but they did not differ between soil sources ( $P \geq 0.08$ ) or with the interaction of debris mass and soil source ( $P \geq 0.16$ ). Linear effects of debris mass were significant for root biomass ( $P < 0.001$ ), while both linear and quadratic effects were significant for shoot biomass ( $P \leq 0.006$ ). Cubic and quadratic regression models provided the best fits for root and shoot biomass, respectively (Table 2). Predicted root biomass decreased 84% as simulated debris mass increased from 1 to  $2 \text{ kg m}^{-2}$  (i.e., from light to heavy debris levels), while predicted shoot biomass decreased only 29% (Figures 5C and D). Because both PPFD and R/FR declined with increasing mass of logging debris, it is not possible to determine which light factor, quantity or quality, caused the reductions in root and shoot biomass of *C. scoparius*. The meta-analysis by Poorter et al. (2012) supported a general finding for terrestrial plants of decreasing allocation to root biomass and increasing allocation to leaves with decreasing light quantity. However, Poorter et al. (2012) also found increased allocation to stem as R/FR decreased below 0.9. Because aboveground biomass of *C. scoparius* was not partitioned into stems and leaves in the current study, it is not possible to determine which light factor, quantity or quality, is most responsible for reduced shoot biomass under high levels of logging debris.

#### Predicting Temperature Limitations to Germination

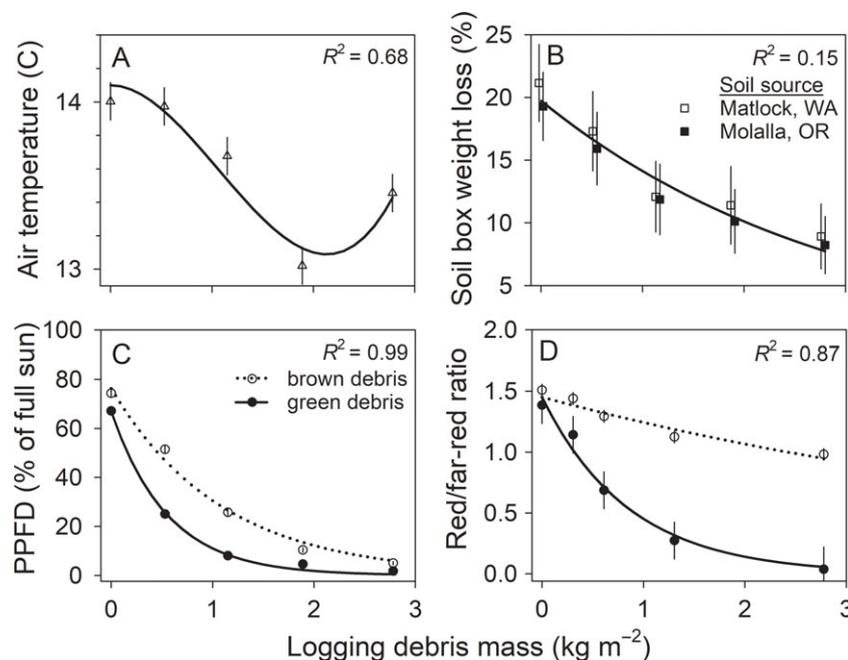
The observed 0.5 C reduction in air temperature for heavy versus light debris (Figure 4A) is predicted to reduce *C. scoparius* germination by less than 1% (Figure 6). If we assume a 1.5 C reduction in temperature for heavy versus light debris, as observed for soil temperature by Harrington et al. (2018), predicted *C. scoparius*

**Table 2.** Regression models characterizing the observed relationships of air temperature in May, weekly soil box weight loss from evaporation, photosynthetic photon flux density (PPFD), red/far-red ratio (R/FR), and *Cytisus scoparius* seedling germination, height, root biomass, and shoot biomass versus mass of simulated logging debris ( $X$ ;  $\text{kg m}^{-2}$ ) in a field study in Olympia, WA, during April to July 2013 (relationships depicted graphically in Figures 4 and 5).

Dependent variable	Regression model <sup>b</sup>	Regression statistics <sup>a</sup>			
		P	$s_{y,x}$	$R^2$	$n$
Air temperature (C)	$Y = 14.1 - 0.668X^2 + 0.209X^3$	<0.001	0.239	0.68	15
Soil box weight loss (%)	$Y = 19.7 \cdot \exp(-0.334X)$	<0.001	9.85	0.15	130
PPFD (% of full sun)	$Y = (76.3 - 9.35G) \cdot \exp[-0.918X - 0.889X(G)]$	<0.001	3.01	0.99	30
R/FR	$Y = 1.47 \cdot \exp[-0.160X - 0.998X(G)]$	<0.001	0.173	0.87	29
Germination (%)	$Y = 31.9$	0.808	20.1	<0.01	15
Height (cm)	$Y = 2.81 - 2.04S + 8.83X - 2.69X^2$	<0.001	1.62	0.64	27
Root biomass (mg)	$Y = 48.3 - 22.9X^2 + 6.06X^3$	<0.001	9.49	0.78	27
Shoot biomass (mg)	$Y = 17.3 - 1.53X^2$	<0.001	4.22	0.52	27

<sup>a</sup>P = significance of the regression;  $s_{y,x}$  = standard error of estimate;  $R^2$  = coefficient of determination adjusted for df; and  $n$  = sample size.

<sup>b</sup>Green indicator variable:  $G = 1$  for green logging debris; and  $G = 0$  for brown logging debris. Soil indicator variable:  $S = 1$  for glacial outwash (Matlock, WA); and  $S = 0$  for cobbly loam (Molalla, OR).



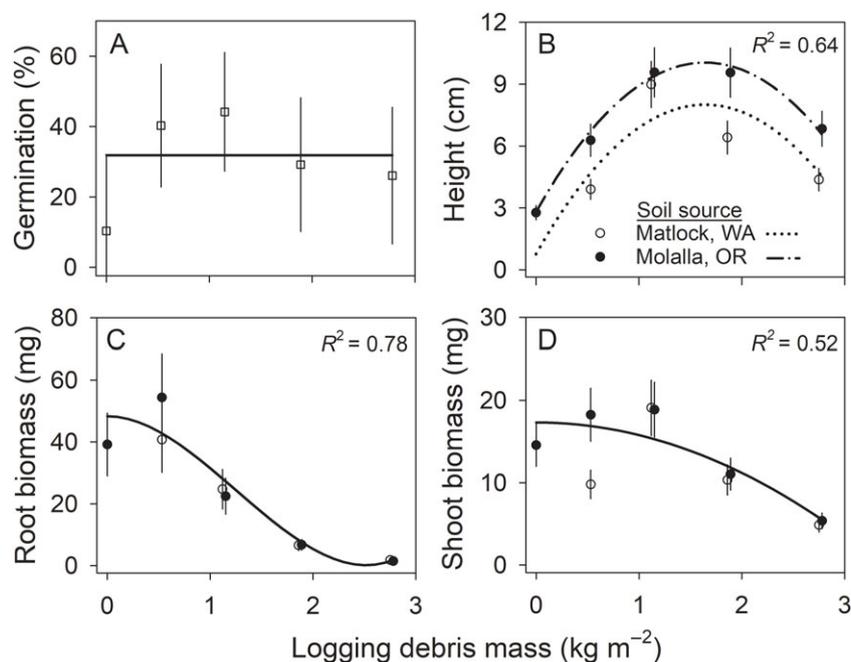
**Figure 4.** Regression relationships of mean values ( $\pm$ SE) of: (A) air temperature in May, (B) weekly soil box weight loss from evaporation, (C) photosynthetic photon flux density (PPFD), and (D) red/far-red light ratio versus varying mass of simulated logging debris in a field study in Olympia, WA, during April to July 2013. Trend lines are from the regression models listed in Table 2.

germination would be reduced by 2.4%. These predictions of potential limitations to germination suggest that temperature is not the primary factor limiting regeneration of *C. scoparius* under heavy debris.

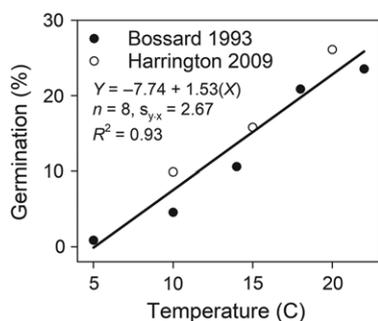
In its native environment, *C. scoparius* does not successfully regenerate without disturbance to the dominant vegetation or soil (Paynter et al. 1998). In a nonnative environment (i.e., California, USA), density of *C. scoparius* germinants declined exponentially during three growing seasons from about 770 to 32 seedlings  $\text{m}^{-2}$ , although the causes of mortality were not identified (Bossard and Rejmánek 1994). In harvested, thinned, and intact forests of coastal Douglas-fir in western Washington, USA, less than 2% of seeds produced seedlings of *C. scoparius* 2 yr after sowing (Harrington 2007). Paynter et al. (1998) suggested that mortality of new seedlings under mature *C. scoparius* in France was attributable to a “rain” of invertebrate herbivores and pathogens;

whereas in open areas, mortality was attributable to competition for light and soil water because new seedlings are relatively poor competitors.

This research has demonstrated that *C. scoparius* seed germination is not responsive to variation in either light quantity or quality. Because of the limited geographic scope of the seed sources used in the studies (i.e., near Matlock, Shelton, and Tumwater, WA), reasonable caution should be used in interpreting these results; seeds from other locations could show variation in their germination and growth responses to the light environment and microclimate. In addition, the simulated logging debris effects study did not assess the effects of physical contact of the debris with the germination substrate or with germinating seedlings, nor did it consider possible debris effects on soil nutrients. Nonetheless, the weak, and potentially exploitable, link in regeneration of *C. scoparius* appears to be not seed germination, but seedling biomass development.



**Figure 5.** Regression relationships of mean values ( $\pm$ SE) of *Cytisus scoparius*: (A) germination, (B) height, (C) root biomass, and (D) shoot biomass versus varying mass of simulated logging debris in a field study in Olympia, WA, during April to July 2013. Seedlings were grown for 90 d in soils collected near Matlock, WA, and Molalla, OR, USA. Trend lines are from the regression models listed in Table 2.



**Figure 6.** Regression relationship of mean germination of *Cytisus scoparius* versus air temperature, based on the laboratory experiments of Bossard (1993) and Harrington (2009). Regression statistics include sample size ( $n$ ), standard error of estimate ( $s_{y,x}$ ), and coefficient of determination ( $R^2$ ).

Based on findings from the laboratory and field studies, reduced biomass of seedlings, especially roots, due to shade of logging debris and subsequently developing native vine and shrub cover (Harrington et al. 2020) creates a significant vulnerability for juvenile *C. scoparius* that makes them susceptible to mortality from summer drought, competition, herbivory, and pathogens. A similar vulnerability was hypothesized for *C. scoparius* seedlings that had reduced root biomass after PRE application of sulfometuron herbicide (Ketchum and Rose 2003). In a 90-d laboratory experiment, Harrington (2009) confirmed the validity of the hypothesis of Ketchum and Rose (2003) by observing a 17% increase in mortality of *C. scoparius* seedlings grown in sulfometuron-treated soil when watering rate was reduced by half midway through the study; nontreated seedlings under the same watering regime had only a 3% to 6% increase in mortality.

Green logging debris was more effective than brown debris at modifying the light environment sufficiently to cause significant reductions in *C. scoparius* root biomass, particularly by reducing R/FR below 1. Forest harvesting in winter or early spring allows green debris to be in place when *C. scoparius* germination begins. Retention of a high level of logging debris after forest harvesting ( $\geq 2 \text{ kg m}^{-2}$ ) has potential application on sites likely to be invaded by *C. scoparius* or those already having the species present in the seedbank.

**Acknowledgments.** Financial support for this research was provided by USDA National Institute for Food and Agriculture (Grants.gov number: GRANT11325729) and USDA Forest Service, Pacific Northwest Research Station. Thanks to James Dollins and Jessyka Williams, PNW Research Station, for technical assistance with the research. No conflicts of interest have been declared.

## References

- Abdullah MM, Jones RA, El-Beltagy AS (1989) An efficient method to overcome seed dormancy in Scotch broom (*Cytisus scoparius*). *Environ Exp Bot* 29:499–505
- Bliss D, Smith H (1985) Penetration of light into soil and its role in the control of seed germination. *Plant Cell Environ* 8:475–483
- Borthwick HA, Hendricks SB, Parker MW, Toole EH, Toole VK (1952) A reversible photoreaction controlling seed germination. *Proc Natl Acad Sci USA* 38:662–666
- Bossard C (2000) *Cytisus scoparius*. Pages 145–149 in Bossard C, Randall J, Hochovsky M, eds. *Invasive Plants of California's Wildlands*. Berkeley: University of California Press
- Bossard CC (1993) Seed germination in the exotic shrub *Cytisus scoparius* (Scotch broom) in California. *Madroño* 40:47–61

- Bossard CC, Rejmánek M (1994) Herbivory, growth, seed production, and resprouting of an exotic invasive shrub, *Cytisus scoparius*. *Biol Conserv* 67:193–200
- Caldwell BA (2006) Effects of invasive Scotch broom on soil properties in a Pacific coastal prairie soil. *Appl Soil Ecol* 32:149–152
- Carter DR, Slesak RA, Harrington TB, Peter DH, D'Amato AW (2019) Scotch broom (*Cytisus scoparius*) alters microenvironment and promotes nonnative grasses. *Biol Invasions* 21:1055–1073
- [CABI] Commonwealth Agricultural Bureau International (2020) Invasive Species Compendium: *Cytisus scoparius* (Scotch Broom). <https://www.cabi.org/isc/datasheet/17610>. Accessed: March 20, 2020
- [CAI] Community Attributes, Inc. (2017) Economic Impact of Invasive Species: Direct Costs Estimates and Economic Impacts for Washington State. Seattle, WA: CAI. 46 p <https://invasivespecies.wa.gov/wp-content/uploads/2019/07/EconomicImpactsRpt.pdf>
- Devine WD, Harrington CA (2007) Influence of harvest residues and vegetation on microsite soil and air temperatures in a young conifer plantation. *Agric For Meteorol* 145:125–138
- DiTomaso JM, Brooks ML, Allen EB, Minnich R, Rice PM, Kyser GB (2006) Control of invasive weeds with prescribed burning. *Weed Technol* 20:535–548
- [EDDMapS] Early Detection and Distribution Mapping System (2020) EDDMapS home page. University of Georgia–Center for Invasive Species and Ecosystem Health. <http://www.eddmaps.org>. Accessed: March 20, 2020
- Grove S, Haubensak KA, Parker IM (2012) Direct and indirect effects of allelopathy in the soil legacy of an exotic plant invasion. *Plant Ecol* 213:1869–1882
- Grove S, Parker IM, Haubensak KA (2015) Persistence of a soil legacy following removal of a nitrogen-fixing invader. *Biol Invasions* 17:2621–2631
- Harrington TB (2007) Establishment of Scotch broom seedlings in Douglas-fir forests: effects of overstorey retention level and seedbed type. Pages 37–42 in Harrington TB, Reichard SH, eds., *Meeting the Challenge: Invasive Plants in Pacific Northwest Ecosystems*. Portland, OR: USDA Forest Service, PNW Res. Sta., Gen. Tech. Rep. PNW-GTR-694. 166 p
- Harrington TB (2009) Seed germination and seedling emergence of Scotch broom (*Cytisus scoparius*). *Weed Sci* 57:620–626
- Harrington TB (2014) Synthetic auxin herbicides control germinating Scotch broom (*Cytisus scoparius*). *Weed Technol* 28:435–442
- Harrington TB, Peter DH, Slesak RA (2018) Logging debris and herbicide treatments improve growing conditions for planted Douglas-fir on a droughty forest site invaded by Scotch broom. *For Ecol Manag* 417:31–39
- Harrington TB, Schoenholtz SH (2010) Effects of logging debris treatments on five-year development of competing vegetation and planted Douglas-fir. *Can J For Res* 40:500–510
- Harrington TB, Slesak RA, Dollins JP, Schoenholtz SH, Peter DH (2020) Logging-debris and vegetation-control treatments influence competitive relationships to limit 15-year productivity of coast Douglas-fir in western Washington and Oregon. *For Ecol Manag*, [10.1016/j.foreco.2020.118288](https://doi.org/10.1016/j.foreco.2020.118288)
- Haubensak KA, Parker IM (2004) Soil changes accompanying invasion of the exotic shrub *Cytisus scoparius* in glacial outwash prairies of western Washington [USA]. *Plant Ecol* 175:71–79
- Holt JS (1995) Plant responses to light: a potential tool for weed management. *Weed Sci* 43:474–482
- Hulting A, Neff K, Coombs E, Parker R, Miller G, Burrill LC (2008) Scotch Broom: Biology and Management in the Pacific Northwest *Cytisus scoparius* (L.) Link. PNW 103. Pacific Northwest Extension (Oregon State University, University of Idaho, and Washington State University). 8 p
- Ivens GW (1983) The influence of temperature on germination of gorse (*Ulex europaeus* L.). *Weed Res* 23:207–216
- Keeley JE (1987) Role of fire in seed germination of woody taxa in California chaparral. *Ecology* 68:434–443
- Ketchum JS, Rose R (2003) Preventing establishment of exotic shrubs (*Cytisus scoparius* (L.) Link. and *Cytisus striatus* (Hill)) with soil active herbicides (hexazinone, sulfometuron, and metsulfuron). *New Forest* 25:83–92
- Mack RN (2003) Plant naturalizations and invasions in the eastern United States: 1634–1860. *Ann Mo Bot Gard* 90:77–90
- Mobley L (1954) Scotch broom, a menace to forest, range and agricultural land. *Proc Calif Weed Sci Soc* 6:39–43
- Neter J, Wasserman W, Kutner MH (1989) *Applied Linear Regression Models*. 2nd ed. Homewood, IL: Richard D. Irwin. Pp 271–284
- O'Connell AM, Grove TS, Mendham DS, Rance SJ (2004) Impact of harvest residue management on soil nitrogen dynamics in *Eucalyptus globulus* plantations in south-western Australia. *Soil Biol Biochem* 36:39–48
- Oneto SR, Kyser GB, DiTomaso JM (2010) Efficacy of mechanical and herbicide control methods for Scotch broom (*Cytisus scoparius*) and cost analysis of chemical control options. *Invasive Plant Sci Manag* 3:421–428
- Paynter Q, Downey PO, Sheppard AW (2003) Age structure and growth of the woody legume weed *Cytisus scoparius* in native and exotic habitats: implications for control. *J Appl Ecol* 40:470–480
- Paynter Q, Fowler SV, Memmott J, Sheppard AW (1998) Factors affecting the establishment of *Cytisus scoparius* in southern France: implications for managing both native and exotic populations. *J Appl Ecol* 35:582–595
- Peter DH, Harrington TB (2018) Effects of forest harvesting, logging debris, and herbicides on the composition, diversity and assembly of a western Washington, USA plant community. *For Ecol Manag* 417:18–30
- Peterson DJ, Prasad R (1998) The biology of Canadian weeds 109 *Cytisus scoparius* (L.) Link. *Can J Plant Sci* 78:497–504
- Poorter H, Niklas KJ, Reich PB, Oleksyn J, Poot P, Mommer L (2012) Biomass allocation to leaves, stems and roots: meta-analyses of interspecific variation and environmental control. *New Phytol* 193:30–50
- Rivas M, Reyes O, Casal M (2006) Influence of heat and smoke treatments on the germination of six leguminous shrubby species. *Int J Wildland Fire* 15:73–80
- Roberts SD, Harrington CA, Terry TA (2005) Harvest residue and competing vegetation affect soil moisture, soil temperature, N availability, and Douglas-fir seedling growth. *For Ecol Manag* 205:333–350
- Rolston MP, Talbot J (1980) Soil temperatures and regrowth of gorse burnt after treatment with herbicides. *New Zeal J Exp Agr* 8:55–61
- Shaben J, Myers JH (2010) Relationships between Scotch broom (*Cytisus scoparius*), soil nutrients, and plant diversity in the Garry oak savannah ecosystem. *Plant Ecol* 207:81–91
- Sheppard, AW, Hodge P, Paynter Q, Rees M (2002) Factors affecting invasion and persistence of broom *Cytisus scoparius* in Australia. *J Appl Ecol* 39:721–734
- Slesak RA, Harrington TB, D'Amato AW (2016a) Invasive Scotch broom alters soil chemical properties in Douglas-fir forests of the Pacific Northwest, USA. *Plant Soil* 398:281–289
- Slesak RA, Harrington TB, Peter DH, DeBruler DG, Schoenholtz SH, Strahm BD (2016b) Effects of intensive management practices on 10-year Douglas-fir growth, soil nutrient pools, and vegetation communities in the Pacific Northwest, USA. *For Ecol Manag* 365:22–33
- Smith H (1982) Light quality, photoperception, and plant strategy. *Annu Rev Plant Physiol* 33:481–518
- Smith JMB, Harlen RL (1991) Preliminary observations on the seed dynamics of broom (*Cytisus scoparius*) at Barrington Tops, New South Wales. *Plant Prot Q* 6:73–78
- Sokal RR, Rohlf JF (1981) *Biometry*. 2nd ed. San Francisco: Freeman. Pp 419–421, 427–428
- [USDA-NRCS] U.S. Department of Agriculture–Natural Resources Conservation Service (2019) Official Soil Series Descriptions. <https://soilseries.sc.egov.usda.gov/osdname.aspx> Accessed: July 22, 2019
- Watt MS, Clinton PW, Whitehead D, Richardson B, Mason EG, Leckie AC (2003a) Above-ground biomass accumulation and nitrogen fixation of broom (*Cytisus scoparius* L.) growing with juvenile *Pinus radiata* on a dryland site. *For Ecol Manag* 184:93–104
- Watt MS, Whitehead D, Mason EG, Richardson B, Kimberly MO (2003b) The influence of weed competition for light and water on growth and dry matter partitioning of young *Pinus radiata*, at a dryland site. *For Ecol Manag* 183:363–376
- Wearne LJ, Morgan JW (2004) Community-level changes in Australian subalpine vegetation following invasion by the non-native shrub *Cytisus scoparius*. *J Veg Sci* 15:595–604
- Williams, PA (1981) Aspects of the ecology of broom (*Cytisus scoparius*) in Canterbury, New Zealand. *New Zeal J Bot* 19:31–43
- Young JA, Young CG (1992) *Seeds of Woody Plants in North America*. Portland, OR: Dioscorides Press. P 131